

1 **Supplementary Materials**

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3 **Bioaccumulation, transfer, and impacts of microplastics in aquatic food chains**

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14 **Plastic Morphology Metric Calculations**

15 To calculate the total volume of plastic in each treatment, we combined 0.1 g (HDPE =
16 0.95 g/cm³) of plastic with 49.9 g of distilled water. This resulted in 0.095 cm³ of
17 plastic in 50 mL of solution. This number was then divided by the total volume
18 estimated for a single plastic particle given its size ($V = 4/3 \times \pi \times r^3$), where $r = 12.5,$
19 35, and 275 μm for the tween, small, and large spheres, respectively. This provided the
20 total number of each type of microplastic sphere present in our stock solutions.

21
22
$$0.095 / (4/3 \times \pi \times r^3) \quad (1)$$

23
24 We multiplied the number of spheres in 0.1 g in 50 mL solution by their surface area (4
25 $\times \pi \times r^2$) to estimate the plastic surface area in each stock solution. Stock solutions were
26 then used to create experimental stocks, that would be added to our jars and tanks,
27 depending on the experiment. We diluted 2% of the stock liquid (1 mL) to 10 mL (9 mL
28 of distilled water). These experimental solutions were then added to experimental
29 systems - for trials carried out in algal/copepod trials, we added 1.5 mL of experimental



30 solution to 148.5 mL of distilled or seawater (150 mL total volume); and for fish trials,
31 we added 10 mL of experimental solution to 1.99 L of freshwater.

32

33 With these estimates of plastic abundance, volume, and surface area, we could then
34 determine the concentration and total surface areas that were contained in each
35 experimental replicate. To estimate the total number of plastics in a jar, we divide the
36 plastics in the 10 mL experimental stock (Equation 1) by 10 and multiply by 1.5. This
37 number is then multiplied by the surface area of a particle to derive the surface area in a
38 jar. To estimate the total number of plastic particles and their surface area in tanks, we
39 divided the numbers present in the experimental solutions (10 mL) by 2000.

40

41 For jars with large spheres:

42 Diameter = 0.055 cm; radius = 0.028 cm; surface area of 1 particle = 0.009 cm²,
43 volume of 1 particle = 0.00009 cm³; plastics in a jar 16.36 particles and 0.155 cm²,
44 plastics in a tank 109 particles and 1.03 cm².

45

46 For jars with small spheres:

47 Diameter = 0.007 cm; radius = 0.0035 cm; surface area of 1 particle = 0.0002 cm²,
48 volume of 1 particle = 1.8×10^{-7} cm³; plastics in a jar 7,939 particles and 1.22 cm²,
49 plastics in a tank 52,924 particles and 8.14 cm².

50

51 **Tween Toxicity Trials**

52 We used fluorescently labeled high-density polyethylene microspheres (“polyspheres,”
53 Cospheric, Santa Barbara, CA, USA) with consistent diameters and density in our
54 exposure trials of algae, copepod consumers, and fish second-order consumers. In
55 accordance with manufacturer recommendations^[1], we understood that a surfactant
56 would promote the suspension of the microspheres in seawater and freshwater media
57 rather than accumulating on the liquid’s surface. Surfactants occur naturally in the
58 environment (Alkyl sulfate, Alcohol ethoxylate, Benzalkonium chloride, *etc.*), and
59 undoubtedly perform the same function in suspending materials in water where the
60 density is similar to the aqueous solution in which they are mixed^[2,3]. The
61 recommended surfactant for this application was Tween 80^[1], and before assessing the
62 potential toxicity of plastics in food webs, we first needed to confirm that Tween80

63 would not impact algal productivity or copepod survival.

64

65 To assess for potential of Tween80 toxicity, we performed exposures of algae and
66 copepods to four experimental treatments: algae grown with the addition of Tween80 (*T*:
67 Tween80 in seawater), microplastics without surfactant (*M*: polyspheres in seawater),
68 Tween & microplastics (*T&M*: Tween80, and polyspheres in seawater), and controls (*C*:
69 seawater only). We used small-bodied marine copepods (*Apocyclops panamensis*) as
70 the primary consumers and the smallest size class of polyspheres (25 μm diameter -
71 largest abundance and SA:Vol). We adjusted the water salinity twice a week to ensure it
72 remained within the optimum range of the copepod consumer (25-30 psu^[4]).

73

74 We suspended fluorescently labeled polyspheres in a concentrated stock solution of: 0.1
75 g of microplastic in 50 g of water. Tween treatment (*T*) stock solutions were comprised
76 of 0.1 g of Tween80 in 50 g of water. Finally, a concentrated stock solution of plastic
77 and Tween (*T&M*) contained 0.1 g of microplastics and 0.1 g of Tween in 50 g of water.
78 All stock solutions were maintained in 50 mL centrifuge tubes away from ambient light
79 to reduce the likelihood of algal growth and light exhaustion of the fluorescent
80 additives.

81

82 Tween 80 exposures did not impact Chlorophyll *a* concentrations ($P = 0.519$, $R^2 =$
83 0.013) or survivorship of, *Apocyclops panamensis* ($P = 0.443$, $R^2 = 0.018$). We
84 therefore proceeded with all subsequent copepod and fish feeding trials with Tween80
85 in all exposure solutions. The use of surfactant in this study did not impact algal growth
86 or copepod survivorship, so we found that Tween80 was a safe means to assess for
87 impacts of plastic exposure to algae and copepods.

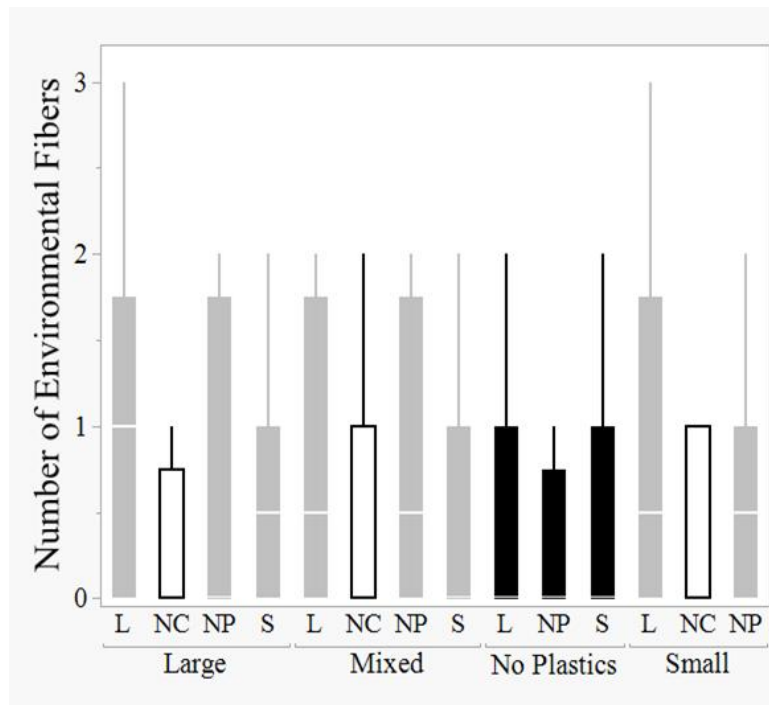
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89 **Environmental Fibers**

90 In our experimental systems, we observed microfibers of unknown origin in addition to
91 the polyspheres and plastic microfibers that we introduced. This was not unexpected, as
92 environmental fibers are often present in natural systems^[5,6]. However, we were able to
93 distinguish them from our experimental plastics based on their shape and lack of
94 fluorescence excitation response. Although we took precautions to prevent
95 cross-contamination, we could not exclude the possibility of airborne or waterborne

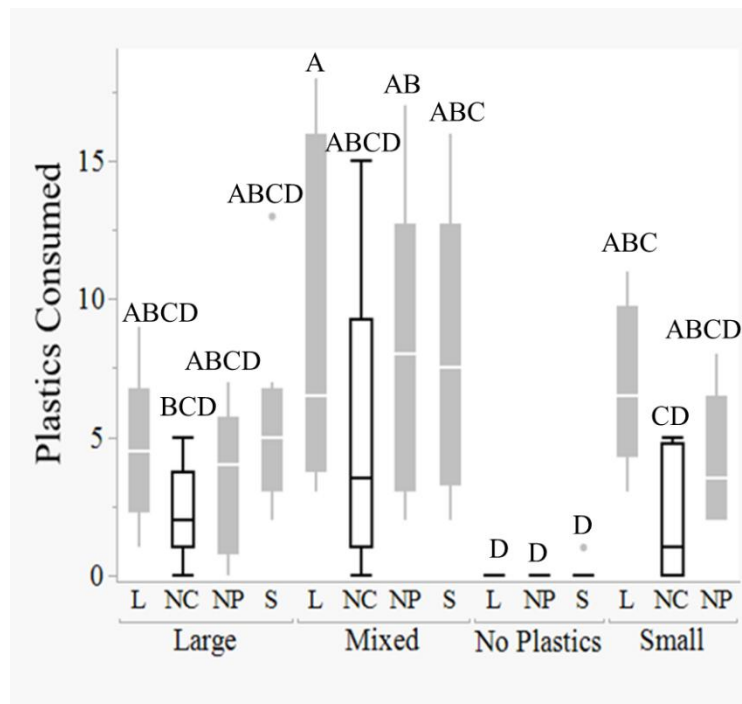
96 contaminants in our facilities. These fibers were consistent across all treatments and did
 97 not appear to be more prevalent in any specific situation [Supplementary Figure 1]. As
 98 a result, we did not include them in our analysis of plastic contaminants. It is worth
 99 noting that such contaminants are commonly reported in natural systems^[7,8].

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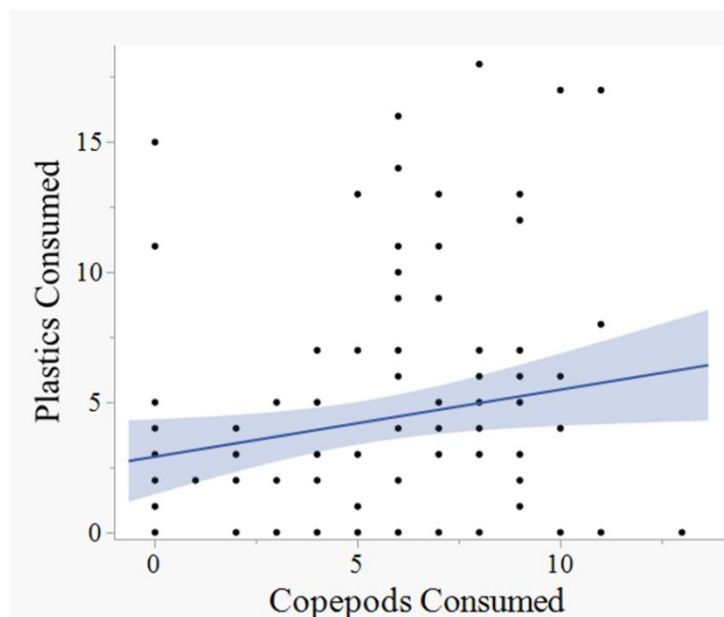


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102 **Supplementary Figure 1.** The number of environmental fibers found during fish
 103 dissections did not vary across trial types. Feeding trials included various types of
 104 plastic exposure to both the fish and the copepods they were consuming. There were
 105 three different types of copepod exposure: large polyspheres (L), small polyspheres (S),
 106 and no plastics (NP). Several trials did not contain copepods (NC). There were also
 107 four types of fish exposure: large polyspheres, large and small polyspheres (Mixed), no
 108 plastics, and small polyspheres. Combinations of exposures resulted in 14 trial types.
 109 Dark filled boxes represent trials where only copepods were available to the fish, light
 110 filled boxes represent trials where fish had both copepods and polyspheres, and open
 111 boxes represent trials where only plastics were given to fish.



112
 113 **Supplementary Figure 2.** There is no discernible trend for the number of plastics
 114 consumed across the fish feeding trials regardless of the presence of plastics (large [L],
 115 small [S], mixed) in copepod cultures or offered as food and copepods without any
 116 plastics exposed (NP). Lettering above the boxes indicates significant differences based
 117 on an alpha value of 0.05 using Tukey’s HSD pairwise post hoc analyses.
 118



119
 120 **Supplementary Figure 3.** The number of experimental plastics (large and small HDPE
 121 microspheres combined) found in the gut of the fish during dissections increased as
 122 more copepods were consumed ($P = 0.0291$, $R^2 = 0.043$).

123 **Estimations for the Preferential Feeding of Fish**

124 In an effort to understand the relative feeding preferences of blacknose dace for
125 copepods or plastics, we conducted a series of trials in which we calculated the
126 expected frequency of encountering each food option. We based our calculations on
127 estimations of the oxygen concentration in the tanks and the respiration rate of the fish.
128 While we acknowledge the roughness of our estimates, we believe they are informative
129 in demonstrating the magnitude of the impact of microplastics on fish feeding behavior.
130 Based on our calculations, we estimate that the oxygen concentration in the tanks was
131 around 8.1 ppm (mg/L) in water^[9]. This indicates that there are approximately 16.2 mg
132 O₂ in the 2L of water in our experimental feeding tanks. According to the general
133 respiration rate of fishes^[10], a fish weighing 5g (0.005 kg, which we used as an
134 approximation for our sample fish population) would require 0.4285 mL O₂ per hour
135 and 0.214 mL O₂ during our 30 min trials. Using the dissolved state at 8.1 mg/L, we
136 estimated that 37.85 mL of water would need to be passed over the gills to acquire the
137 required oxygen.

138

139 If fish consumed plastics at a rate that was close to random encounter, and assuming
140 that the fish and plastics maintained their position in the tank (which we acknowledge
141 is an unrealistic assumption), we would expect that only 1,001 of the 52,900
142 polyspheres in the tank would be present in the gut of the fish for small plastics (70 mm
143 diameter). Similarly, we would expect fewer only one of the 109 large (550 mm)
144 plastics to be consumed if they were ingested consistent with their encounter rate.

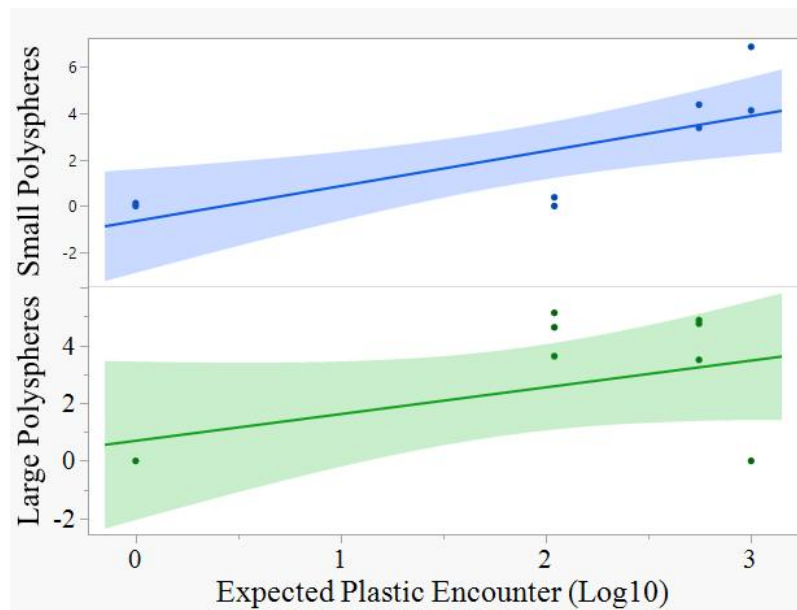
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146 Lastly, we can estimate the random encounter rate for a copepod with a fish predator by
147 dividing the number of invertebrates ($n = 15$) by the tank volume (2,000 mL) and the
148 proportion of the tank that the fish will breathe (37.85/2,000 mL). Based on this
149 calculation, we can expect fewer than 1 copepod will encounter a fish in each 30 min
150 trial. While these estimates are conservative in nature given the ability of copepods to
151 move in the water and likely movement of fish in the tank, they allow for a baseline
152 null with which to compare our experimental data.

153

154 The consumption rates of plastics were lower than expected for small (70 mm)
155 polyspheres (95% CI = 5.54-2.76, DF = 1, Chi = 2,523, $P < 0.01$) but not for large (550

156 mm) polyspheres (95% CI = 1.72-3.10, DF = 1, Chi = 1.60, $P > 0.05$). The
 157 consumption rates of plastics [Supplementary Figure 4] were lower than expected for
 158 small (70 mm) polyspheres (95% CI = 5.54-2.76, DF = 1, Chi = 2,523, $P < 0.01$) but
 159 not for large (550 mm) polyspheres (95% CI = 1.72-3.10, DF = 1, Chi = 1.60, $P > 0.05$).
 160 The consumption rates for copepods was greater than expected (95% CI = 5.95-7.07,
 161 DF = 2, Chi = 14.66, $P < 0.01$).
 162



163
 164 **Supplementary Figure 4.** The number of small plastics (70mm, top) consumed by fish
 165 in feeding trials increased as the density of plastics increased; however, the rate was
 166 still lower than expected by chance ($F = 65.3$, $R^2 = 0.879$, estimate = 0.006, $P <$
 167 0.0001). The number of large (550 mm, bottom) also consistently increased with the
 168 frequency of polyspheres available for consumption ($F = 934.1$, $R^2 = 0.791$, estimate =
 169 0.044 , $P = 0.000$).
 170

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